



### **Jacobs Journal of Environmental Sciences**

Research Article

# Effects of Mixed-Grazing on Greenhouse Gas Fluxes During the Growing Season in Desert Steppe

Dan Ding<sup>1</sup>, Andreas Wilkes<sup>2</sup>, Xiaojuan Liu<sup>1</sup>, Qing Chen<sup>3</sup>, Chengjie Wang<sup>\*</sup>

<sup>1</sup>College of Ecology and Environmental Science, Inner Mongolia Agricultural University, Hohhot 010019, Inner Mongolia, China

<sup>2</sup>Values for Development Limited, Bury St Edmunds, IP333EQ, UK

<sup>3</sup> Tianjin Key Lab of Water Resources and Environment, Tianjin Normal University, Tianjin 300387, China

\*Corresponding author: Chengjie Wang, College of Ecology and Environmental Science, Inner Mongolia Agricultural University,

Hohhot 010019, Inner Mongolia, China, Tel: +8615849162125; Email: nmgcjwang3@163.com

Received: 08-27-2015

Accepted: 01-15-2016

Published: 03-24-2016

Copyright: © 2016 Chengjie Wang

#### **Abstract**

We studied the effects of different grazing species (cattle grazing (CG), sheep grazing (SG) and mixed-grazing of cattle and sheep (MG)) on fluxes of carbon dioxide ( $\rm CO_2$ ) and methane ( $\rm CH_4$ ) during the growing season in a desert steppe of Inner Mongolia, China. The static chamber method was used to measure fluxes of  $\rm CO_2$  and  $\rm CH_4$  from June to October, 2013. Results indicate that the grazed desert steppe was a net source of soil atmospheric  $\rm CO_2$  exchange and a net sink of soil atmospheric  $\rm CH_4$  exchange during the growing season. Grazing species did not alter the  $\rm CO_2$  or  $\rm CH_4$  flux direction.  $\rm CO_2$  flux in the MG plot (256.66 mg m<sup>-2</sup> h<sup>-1</sup>) was lower than that in CG (351.18 mg m<sup>-2</sup>h<sup>-1</sup>) and SG (315.38 mg m<sup>-2</sup> h<sup>-1</sup>) plots.  $\rm CH_4$  flux in the MG plot (-0.1330 mg m<sup>-2</sup> h<sup>-1</sup>) was higher than in the CG (-0.1120 mg m<sup>-2</sup> h<sup>-1</sup>) and CK (-0.1099 mg m<sup>-2</sup> h<sup>-1</sup>) plots. Binomial regression equations of the fluxes of  $\rm CO_2$  ( $\rm R^2 = 0.509$ ) and  $\rm CH_4$  ( $\rm R^2 = 0.327$ ) on soil temperature and moisture were developed. These findings imply that MG reduced  $\rm CO_2$  emission by 26.9 % and 18.6 %, and increased  $\rm CH_4$  uptake by 18.8 % and 14.7 % compared with CG and SG, respectively, in a desert steppe of Inner Mongolia.

Keywords: Greenhouse Gas Flux; Mixed-Grazing; Growing Season; Inner Mongolia Desert Steppe

#### **Abbreviations:**

CG: Cattle Grazing; SG: Sheep Grazing;

MG: Mixed-Grazing of Cattle and Sheep;

GHG: Greenhouse Gas

#### Introduction

Soil-atmosphere greenhouse gas (GHG) exchange in terrestrial ecosystems has an important effect on global climate change. The carbon dioxide ( $\rm CO_2$ ) concentration in the atmosphere has reached 400 ppm and the average global temperature has increased by 0.85 °C from 1880 to 2012 [1]. Although methane ( $\rm CH_4$ ) flux from the soil-atmosphere is usually smaller than  $\rm CO_2$  flux, the global warming poten-

tial on a molar basis of  $\mathrm{CH_4}$  over a 100-year timeframe is 34 times greater than  $\mathrm{CO_2}$  [2]. Factors affecting the sources and sinks of GHGs in terrestrial ecosystems are therefore an important topic of study. Several studies of  $\mathrm{CO_2}$  and  $\mathrm{CH_4}$  fluxes from soils have focused on agricultural and forest ecosystems in the temperate zones of Europe and North America [3-5], and there are more limited reports for semi-arid grassland ecosystems in the Eurasian steppe [6-8].

Approximately 28 % of the Eurasian steppe and 91 % of the North American prairies (the two historically dominant temperate grassland ecosystems) have already been converted to arable land or other land uses [9]. Livestock grazing is the main land use for remaining steppe and prairie. Over a quarter of the global potential for soil C storage may be influenced by grazing [10]. Direct grazer impact on plant production, and thereby on potential soil C inputs, is likely mediated by grazing intensity, and has been extensively studied by means of theoretical models [11] and through experimentation [12]. Grazers can also indirectly alter plant community composition through their diet selectivity [13,14], and consequently influence soil C inputs [15].

Animal excreta is an important source of GHG [16-19]. Sheep and cattle excreta patches are different in nature by the area covered, nutrient concentration in soil, and the height of fall of excreta [20]. It has been reported that in Inner Mongolian grassland grazing changes soil moisture holding capacity, which in turn affects GHG fluxes [8,21,22]. [23] found that soil water content and temperature were the main factors driving GHG fluxes. Other studies reported that grazing changes the community of soil methanotrophs in typical steppe [24]. [24] found that population of methanotrophs in the topsoil of desert steppe was higher under light grazing and moderate grazing sites than in non-grazed and heavily grazed sites. In tropical conditions, [25] reported that the average biomass of the experimental area was lower with mixed grazing animals (cattle and goats) compared with goat grazing alone, but the average daily gain of goats in mixed grazing conditions was higher than goat grazing alone, suggesting better use of the sward. Grazer type has also been found to influence some characteristics of alfalfa-orchard grass pasture, with mixed grazing promoting a more homogenous harvest than grazing with heifer only, due to lower rejection of dung-contaminated forage [26]. Therefore, the amount of soil C storage may be higher than other grazing species, for example cattle and sheep grazing.

Previous studies on mixed grazing have seldom researched soil-atmosphere GHG exchange in the study region. Therefore, our aims here were to 1) assess the effects of the different grazer species (sheep grazing, cattle grazing, and cattle and sheep mixed-grazing) on CO<sub>2</sub> and CH<sub>4</sub> fluxes during the growing season in an Inner Mongolian desert steppe; and 2) analyze the relationship between GHG exchange and environmental parameters (soil temperature and moisture). We hypothesized that 1) the different intake behaviors of cattle and sheep grazing affect vegetative characteristics, which in turn affects soil properties and respiration, and 2) soil moisture content and temperature are the principal factors controlling the fluxes of CO<sub>2</sub> and CH<sub>4</sub>.

#### **Materials and Methods**

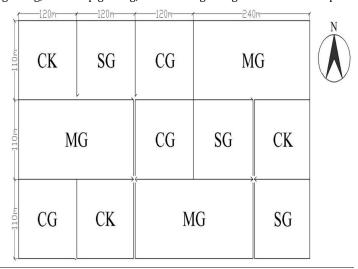
#### Study site

The study was conducted in Xilamuren desert steppe, located at 41° 21′ N, 111° 112′ E, at an elevation of approximately 1,602 m above sea level. The regional climate is temperate continental, characterized by a short growing season and long cold winter, with a frost-free period of 83 d. The average annual precipitation is approximately 284 mm, of which nearly 75 % falls from July to September, and evaporation is 2305 mm. The average annual temperature is 2.5 °C and there are 3100–3300 sunshine hours. The dominant soil types are Kastanozem (FAO soil classification) or Brown Chernozem (Canadian Soil Classification) with a loamy sand texture. The steppe is dominated by *Stipa krylovii Griseb, Artemisia frigida* Willd, and accompanied by *Leymus chinensis Tzvel, Stellera chamaejasme* Linn, *Leynus secalinus* Tzvel, and *Heteropappus altaicus* Novopokr.

#### **Experimental design**

The fluxes of CO<sub>2</sub> and CH<sub>4</sub> were measured in twelve grazing plots in 2013. Before a grazing experiment began on these plots in 2012, the area had been under free grazing. The grazing experiment that started in 2012 was composed of four treatments of cattle grazing (CG, 3 cattle per plot), sheep grazing (SG, 15 sheep per plot), mixed-grazing of cattle and sheep (MG, 3 cattle and 15 sheep per plot), and no grazing (CK) with three replicates for each treatment (Figure. 1). Each plot was grazed in the first seven days from June to September only. During this period, sheep and cattle grazed 24 h per day. Then sheep and cattle were moved to other grazing areas after 7 days of grazing. The grazed sheep and cattle were all 1.5 years old. The stocking rate was considered a moderate grazing intensity based on livestock intake, species composition and ground cover in the growing season.

**Figure.1** Schematic diagram of the experimental area. CG: cattle grazing; SG: sheep grazing; MG: mixed-grazing of cattle and sheep



#### Measurements of GHGs and environmental factors

Fluxes of CH<sub>4</sub> and CO<sub>2</sub> were measured using the closed static chamber method [27]. The chamber had a dimension of 50 ×  $50 \times 50$  (cm) made of stainless steel. A 9 VDC fan was fixed to the top wall of each chamber to mix the chamber atmosphere. The chamber was covered with a shroud made of camel hair, aluminum foil and white canvas to limit heating of the chamber atmosphere during sampling. Three points were randomly selected in each plot. At each point, the chamber was placed on a steel base frame driven 10 cm into the soil one month prior to the start of the experiment. The base frame had a channel in which the chamber was inserted and the channel was filled with water to seal the chamber atmosphere. During gas flux determination, a disposable syringe (100 ml) with a 3-way valve was used to collect 200 ml of chamber atmosphere into a sample gas bag (Dalian Hede Technologies Co., Ltd., Dalian, China) at 10 min intervals over a 30 min period. The gas samples were taken between 9:30-10:30 am, which are times that are representative of the average rate of CH<sub>4</sub> and CO<sub>2</sub> fluxes over a 24-h cycle [28]. Soil temperature and moisture (0-10 cm) were measured by thermocouples, a hand-held reader (HH-25TC, OMEGA Engineering Inc., Stamford, CT) and a portable TDR probe (HH2, Delta-T Devices, Cambridge, UK) at the same time as the measurement of GHG fluxes. The concentrations of CO<sub>2</sub> and CH<sub>4</sub> in the gas samples were analyzed using a cavity ring-down spectrophotometer (Picarro G1301, Santa Clara, USA). Gas fluxes were calculated using the following equation:

$$F = \rho \cdot \frac{V}{A} \cdot \frac{\Delta c}{\Delta t} \tag{1}$$

where F is the flux (mg m<sup>-2</sup>2 h<sup>-1</sup>) of  $CO_2$  or  $CH_4$ ;  $\rho$  is the density of 1 mol  $CO_2$  or  $CH_4$  (kg m<sup>-3</sup>);  $\Delta c \Delta t^{-1}$  is the rate of change in gas concentration (h<sup>-1</sup>); V and A are the volume (m<sup>3</sup>) and the chamber base area (m<sup>2</sup>), respectively.

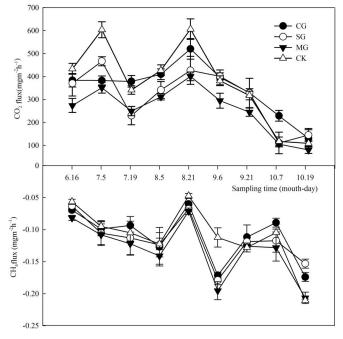
#### **Statistical Analysis**

The fluxes of CO<sub>2</sub> and CH<sub>4</sub> were analyzed using the MIXED procedure of the Statistical Package for Social Science (SPSS 13.0 for Windows, 2003), to test the significance of differences in CO2 and CH4 fluxes. Replicate flux measurements were averaged over sampling points for each grazing plot. Grazer species, period and their interactions were treated as fixed effects, with plot as a random effect, and sampling date as the repeated measure with the grazing plot used as the subject. The data were examined for homogeneity of variances and for normal distribution before analysis. Where necessary, data were transformed using log transformation. Paired means of significant differences in treatments were determined using the least significant difference (LSD) statistic. To test the correlations between soil temperature moisture and the fluxes of CO<sub>2</sub> and CH<sub>4</sub>, Pearson's correlation analysis was performed. The R<sup>2</sup> (square of Pearson correlation coefficient) value was used to identify the best fit function (i.e. linear or quadratic). All significances mentioned in the paper are at the p = 0.05 level unless otherwise noted.

#### Results

#### Effect of grazing system on GHG fluxes

The dynamic feature of  $\mathrm{CO}_2$  fluxes for CG, SG, MG and CK are shown in Figure. 2. The measured mean  $\mathrm{CO}_2$  fluxes for CG, SG, MG and CK were 351.18, 315.38, 256.66 and 369.96 mg m<sup>-2</sup> h<sup>-1</sup>, respectively, which indicates that the experimental plots were a net source of soil-atmospheric  $\mathrm{CO}_2$  exchange during the entire measurement period (Table 1).  $\mathrm{CO}_2$  emission occurred mainly in August.



**Figure 2.** Seasonal variation of carbon dioxide  $(CO_2)$  and methane  $(CH_4)$  fluxes in different grazing species CG: cattle grazing; SG: sheep grazing; MG: mixed-grazing of cattle and sheep

The measured mean  $\mathrm{CH_4}$  fluxes for CG, SG, MG and CK were -0.112, -0.116, -0.133 and -0.110 mg m $^{-2}$  h $^{-1}$ , respectively, which indicates that the experimental plots were a net sink of soil-atmospheric  $\mathrm{CH_4}$  exchange during the entire measurement period.

The pure carbon (C) flux converted from flux of  $\rm CO_2$  and  $\rm CH_4$  is shown in Table 1. The measured mean C flux for CG, SG, MG and CK were 348.87, 315.27, 262.49 and 367.01 mg m $^{-2}$  h $^{-1}$ , respectively, which indicates that the grazed desert steppe is a net source of soil-atmospheric C exchange during the growing season. There was lower C emission in the MG plot compared with CG, SG and CK plots during the growing season.

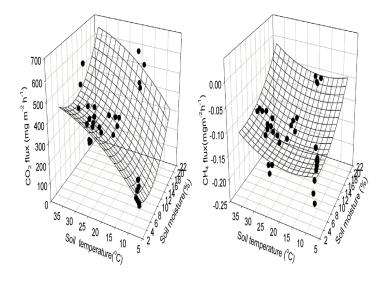
<b>Table 1.</b> Comparison of greenhouse gas emission in different grazing
--

Tuble 1. doinput ison of greenhouse gas emission in affecting species.												
Item	Treatment											
	CG	SG	MG	CK	SEM	P value						
$CO_2$ flux (mg m <sup>-2</sup> h <sup>-1</sup> )	351.17±14.51	315.38±15.73	256.66±15.40	369.96±25.63	9.452	0.967						
$CH_4$ flux $(mg m^{-2} h^{-1})$	0.112±0.006	0.116±0.005	0.133±0.007	0.110±0.007	0.003	0.745						
C flux $(\text{mg m}^{-2} \text{ h}^{-1})$	348.87±15.30	315.27±15.73	262.49±15.57	367.01±25.80	9.594	0.891						

CG: cattle grazing; SG: sheep grazing; MG: mixed-grazing of cattle and sheep; CK: non-grazing; SEM: standard error of mean; P: probability.  $F_{co2}$ -C + FCH<sub>4</sub>-C = F•MC /M<sub>CO2</sub> + F•MC /MCH<sub>4</sub>

## Relationship between GHG fluxes and environmental factors

The relationship between soil temperature and moisture and the measured fluxes  $\mathrm{CO}_2$  and  $\mathrm{CH}_4$  under the four grazing treatments are shown in Figure. 3. The results indicate that  $\mathrm{CO}_2$  flux was positively correlated with topsoil (0–10 cm) temperature ( $\mathrm{R}^2$  = 0.540, p < 0.01) and moisture ( $\mathrm{R}^2$  = 0.523, p < 0.01).  $\mathrm{CH}_4$  flux was negatively correlated ( $\mathrm{R}^2$  = 0.532, p < 0.01) with topsoil moisture, whereas no significant relationship between  $\mathrm{CH}_4$  flux and topsoil temperature was observed ( $\mathrm{R}^2$  = 0.162, p > 0.05) in this study. Considering the combined effect, binomial regression equations of the fluxes of  $\mathrm{CO}_2$  ( $\mathrm{R}^2$  = 0.509) and  $\mathrm{CH}_4$  ( $\mathrm{R}^2$  = 0.327) on soil temperature and moisture were developed in this study (Table 2).



**Figure 3.** Correlation between carbon dioxide  $(CO_2)$  and methane  $(CH_4)$  fluxes with soil temperature and soil moisture

**Table 2.** Binomial regression equation between  ${\rm CO_2}$  and  ${\rm CH_4}$  fluxes with soil temperature and soil moisture

-	$F = y_0 + aM_s + bT_s + cM_s^2 + dT_s^2$										
Item	<b>y</b> 0	a	b	c	d	n	$R^2$				
CO <sub>2</sub>	70.435	-35.035	24.410	1.870	-0.304	96	0.509				
CH <sub>4</sub>	-0.117	-0.0031	-0.0034	0.0004	0.0003	96	0.327				

Ms: soil moisture; Ts: soil temperature;

#### **Discussion**

Grassland ecosystems play a key role in C and N cycles, and are also sensitive to the impact of grazing [29]. Grazing is a complex event and improper grazing practices have led to grassland degradation in Inner Mongolia, China. Grassland degradation has a significant feedback to biosphere-atmosphere C exchange because grasslands can serve either as a source or sink. Overgrazing reduces below ground net primary production and thereby the C storage capacity and turnover via the volatilization and dislocation by erosion of a huge amount of C previously stored in the ecosystem [30]. Thus, it was urgent to identify suitable grazing practices for management of grassland ecosystems and prevention of ecosystem degradation.

The grazed desert steppe in Inner Mongolia is a net source of soil atmospheric C exchange during the growing season, when measured as the sum of soil and plant respiration. At the same time, grassland soil where methanotrophic activity occurs is an important natural sink for CH<sub>4</sub>. In MG sites, CO<sub>2</sub> flux was lower and CH<sub>4</sub> flux was higher that in the other grazing systems. This finding suggests that planned adjustment of grazer species can reduce CO<sub>2</sub> emission and increase CH<sub>4</sub> sequestration. This maybe caused by glomalin, a secretion of arbuscular mycorrhizal fungi, which the study [31] has shown is an important source of the active soil organic carbon pool. Total glomalin storage under MG has been found to be higher than under grazing by single species (i.e. SG and CG) [32].

Soil carbon storage was higher in grazed area than in ungrazed areas in our study, which is similar to results reported by the study [33], who showed that grazed rangelands in mountain

meadows had a greater proportion of active C in total soil C pools. This is attributed to the effects of grazing on plant biomass, regrowth, community structure, soil biotic, and abiotic factors [34]. Therefore, grazing affects both the physical and chemical properties of soil. The chemical composition and rate of decomposition of plant residues are important determinants of C accumulation in the soil [35]. [36] reported that high grazing intensity reduced CH<sub>4</sub> uptake significantly by 37.9 % compared with non-grazed steppe, while [7] did not detect significant differences in CH, uptake between grazed and nongrazed steppes. This discrepancy in the literature stems from the differences in soil characteristics, degree of degradation associated with grazing, age of exclosures, contribution of manure to the overall nutrient balance in the ecosystem, climatic conditions, and original and post-grazing vegetation communities.

Our study found that the magnitude of GHG fluxes was strongly influenced by soil moisture, as has also been observed in other studies [4,37-39] These studies also reported that soil temperature is an important factor affecting microbial activity. [40] observed that the daily CH, flux and topsoil temperature followed an exponential trend during grazing in a semiarid grassland, with soil temperature explaining 66-82 % of the variation in the daily CH<sub>4</sub> uptake. We did not find a significant correlation between CH, flux and topsoil temperature, which is similar to findings reported by [41,42]. This is ascribed to the effects of soil moisture, as rainfall in the study area is low. [43] reported that CH, oxidation was limited due to low microbial activity at low soil moisture conditions, while it was mainly determined by temperature for wet soils. In our study, there were correlation and regression relationships between soil environmental factors (temperature and moisture) and GHG fluxes.

#### Conclusion

The grazed desert steppe grassland soil was a  $\mathrm{CO_2}$  source and a  $\mathrm{CH_4}$  sink for soil-atmospheric exchange. Mixed grazing species (i.e. cattle and sheep) reduced  $\mathrm{CO_2}$  emission by 26.91 % and 18.62 %, and increased  $\mathrm{CH_4}$  uptake by 18.75 % and 14.66 % compared with CG and SG, respectively. The interaction of soil moisture and temperature was the principal factor controlling soil-atmosphere GHG exchange in the grazed desert steppe in Inner Mongolia, China.

#### Acknowledgements

This research was funded by the National Natural Science Foundation of China (31460125 and 31300386), the Ministry of Education of the P.R. China (213006A), the National Science and Technology Support Program (2012BAD13B00), and the Science and Technology Innovation Group of the Ministry of Education (IRT1259).

#### References

- 1. IPCC Climate Change 2013: Mitigation. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. 2013, University of Berne, Switzerland.
- 2. IPCC Climate change 2014: impacts, adaptation, and vulnerability. 2014, Cambridge University Press.
- 3. Mosier AR, Parton WJ, Valentine DW, Ojima DS, Schimel DS et al.  $CH_4$  and  $N_2O$  fluxes in the Colorado short grass steppe: Impact of landscape and nitrogen addition. Global Biogeochemical Cycles. 1996, 10: 387-399.
- 4. Butterbach-Bahl K. Trace Gas Exchange in Forest Ecosystems. The Netherlands: Kluwer, Dordrecht, pp. 2002, 141-156.
- 5. Groffman PM, Hardy JP, Driscoll CT, Fahey TJ. Snow depth, soil freezing and fluxes of carbon dioxide, nitrous oxide and methane in a northern hard-wood forest. Global Change Biology. 2006, 12: 1748-1760.
- 6. Dong YS, Zhang S, Qi YC, Chen ZZ, Geng YB. Fluxes of  ${\rm CO_2}$ ,  ${\rm N_2O}$  and  ${\rm CH_4}$  from a typical temperate grassland in Inner Mongolia and its daily variation. Chinese Science Bulletin. 2000, 45: 1590-1594.
- 7. Wang YS, Xue M, Zheng XH, Ji BM, Du R. Effects of environmental factors on  $\rm N_2O$  emission from and  $\rm CH_4$  uptake by the typical grasslands in the Inner Mongolia. Chemosphere. 2005, 58: 205-215.
- 8. Liu CY, Holst J, Bruggemann N, Butterbach-Bahl K, Yao ZS et al. Winter grazing reduces methane uptake by soils of a typical semi-arid steppe in Inner Mongolia, China. Atmosphere Environment. 2007, 41(28): 5948-5958.
- 9. White RP, Murray S, Rohweder M. Pilot Analysis of Global Ecosystem: Grassland Ecosystems. Washington: World Resource Institute.2000.
- 10. Scurlock JM, Hall DO. The global carbon sink: a grassland perspective. Global Change Biology.1998, 4(22): 229-233.
- 11. Mazancourt C , Loreau M. Effect of herbivory and plant species replacement on primary production. American Naturalist. 2000, 155: 735-754.
- 12. Milchunas DG, Lauenroth WK. Quantitative effects of grazing on vegetation and soils over a global range of environments. Ecological Monographs.1993, 63(4): 327-366.
- 13. Augustine DJ, McNaughton SJ . Ungulate effects on functional species composition of plant communities: herbivore selec-

- tivity and plant tolerance. The Journal of Wildlife Management. 1998, 62:1164-1182.
- 14. Ritchie ME, Olff H. Herbivore diversity and plant dynamics: additive and compensatory effects. Blackwell Science, Oxford. 1999, 175-204.
- 15. De Deyn GB, Cornelissen JHC, Bardgett RD.Plant functional traits and soil carbon sequestration in contrasting biomes. Ecology Letter. 2008, 11: 516-531.
- 16. Clemens J, Ahlgrimm HJ.Greenhouse gases from animal husbandry: mitigation options. Nutrient Cycling Agroecosystems. 2001, 60: 287-300.
- 17. Bol R, Petersen SO, Christofides C, Dittert K, Hansen MN. Short-term  $N_2O$ ,  $CO_2$ ,  $NH_3$  fluxes, and N/C transfers in a Danish grass-clover pasture after simulated urine deposition in autumn. Journal of Plant Nutrition and Soil Science. 2004, 167: 568-576.
- 18. Saggar S, Bolan NS, Bhandral R, Hedley CB, Luo J. A review of emissions of methane, ammonia, and nitrous oxide from animal excreta deposition and farm effluent application in grazed pastures. New Zeland Journal of Agriculture Respiration. 2004, 47(4): 513-544.
- 19. Cardenas LM, Chadwick D, Scholefield D, Fychan R, Marley CL et al. The effect of diet manipulation on nitrous oxide and methane emission from manure application on incubated grassland soils. Atmosphere Environment. 2007, 41(33): 7096-7107.
- 20. Williams RH, Haynes RJ.Comparison of initial wetting pattern, nutrient concentrations in soil solution and the fate of 15N-labelled urine in sheep and cattle urine patch areas of pasture soil. Plant Soil. 1994, 162: 49-59.
- 21. Qi YC, Dong YS, Yang XH, Geng YB, Liu LX. Effects of grazing on carbon dioxide and methane fluxes in typical temperate grassland in Inner Mongolia, China. Resources Science. 2005, 02: 103-109.
- 22. Wolf B, Zheng XH, Bruggemann N, Chen WW, Dannenmann M et al. Grazing-induced reduction of natural nitrous oxide release from continental steppe. Nature. 2010, 08931: 881-884.
- 23. Simona C, Angelo F. Soil-atmosphere methane exchange in undisturbed and burned Mediterranean shrubland of southern Etaly. Ecosystems. 2005, 8: 182-190.
- 24. Zhou XQ, Wang YF, Huang XZ, Tian JQ. Effect of grazing intensities on the activity and community structure of methane-oxidizing bacteria of grassland soil in Inner Mongolia. Nutrient Cycling in Agroecosystems. 2008, 80: 145-152.

- 25. d'Alexis S, Periacarpin F, Jackson F, Boval M. Mixed grazing systems of goats with cattle in tropical conditions: an alternative to improving animal production in the pasture. Animal. 2014, 8: 1282-1289.
- 26. Mendiola-Gonzalez, Martinez-Hernandez, Cortes-Diaz .Effect of mixed and single grazing on an alfalfa-orchard pasture. Agriculture.2007, 4: 395-403.
- 27. Wang YS, Wang YH. Quick measurement of  $\mathrm{CH_4}$ ,  $\mathrm{CO_2}$  and  $\mathrm{N_2O}$  emission from a short-plant ecosystem. Advances in Atmosphere Science. 2003, 20(5): 842–844.
- 28. Tang XL, Liu SG, Zhou GY, Zhang DQ, Zhuo CY. Soil-atmospheric exchange of  ${\rm CO_2}$ ,  ${\rm CH_4}$ , and  ${\rm N_2O}$  in three subtropical forest ecosystems in southern China. Global Change Biology. 2006, 12(3): 546-560.
- 29. Li X, Chen Z . Soil microbial biomass C and N along a climatic transect in the Mongolian steppe. Biology and Fertility of Soils. 2004, 39:344-351.
- 30. Schlesinger WH, Reynolds JF, Cunningham GL, Huenneke LF, Jarell WM et al. Biological feedbacks in global desertification. Science. 1990, 247(4946): 1043-1048.
- 31. Rillig MC, Rmsey PW, Morris S, Paul EA. Glomalin, an arbuscular-mycorrhizal fungal soil protein, responds to land-use change. Plant Soil. 2003, 253: 293-299.
- 32. Bai HH, Bao XL, Sun XM, Jiang SW. The effect of stocking rate on soil glomalin under traditional and mixed grazing systems in a temperate steppe. Proedia Environmental Sciences. 2011, 11: 817-823.
- 33. Gill R A. Influence of 90 years of protection from grazing on plant and soil processes in the subalpine of the Wasatch Plateau, USA. Range Ecological Management. 2007, 60: 88-98.
- $34.\ Kioko\ J,\ Kiringe\ JW,\ Seno\ SO\ .$  Impacts of livestock grazing on a savanna grassland in Kenya. Journal of Arid Land . 2012, 4(1): 29-35.
- 35. Whalen JK, Willms WD, Dormaar JF. Soil carbon, nitrogen and phosphorus in modified rangeland communities. Journal of Range Management. 2003, 56(6): 665-672.
- 36. Tang SM, Wang CJ, Wilkes A, Zhou P, Jiang YY et al.Contribution of Grazing to Soil-atmosphere  $\mathrm{CH_4}$  Exchange during the Growing Season in a Continental Steppe. Atmospheric Environment. 2013, 67: 170-176.
- 37. Koschorreck M, Conrad R. Oxidation of atmospheric methane in soil: measurements in the field, in soil cores and in soil samples. Global Biogeochemical Cycles. 1993, 7(1): 109-121.

38. Shrestha BM, Sitaula BK, Singh BR, Bajracharya RM . Fluxes of  $\mathrm{CO_2}$  and  $\mathrm{CH_4}$  in soil profiles of a mountainous watershed of Nepal as influenced by land use, temperature, moisture and substrate addition. Nutrient Cycling in Agroecosystems. 2004, 68(2): 155-164.

- 39. Chen YP, Li YQ, Awada T, Han JJ, Luo YQ. Carbon sequestration in the total and light fraction soil organic matter along a chronosequence in grazing exclosures in a semiarid degraded sandy site in China. Journal of Arid Land. 4(4): 411-419.
- 40. Chen W, Wolf B, Zheng X, Yao Z, Butterbach-Bahl K et al. Annual methane uptake by temperate semiarid steppes as regulated by stocking rates, aboveground plant biomass and topsoil air permeability. Global Change Biology. 2011, 17: 2803-2816.

- 41. Del Grosso SJ, Parton WJ, Mosier AR, Ojima DS, Potter CS et al.General CH4 oxidation model and comparisons of CH4 oxidation in natural and managed systems. Global Biogeochemical Cycles. 2000, 14: 999-1019.
- 42. Merino A, Prez-Batall P, Macas F. Responses of soil organic matter and greenhouse gas fluxes to soil management and land use changes in a humid temperate region of southern Europe. Soil Biology and Biochemistry. 2004, 36(6): 917-925.
- 43. Christensen TR. Methane emission from arctic tundra. Biogeochemistry. 1993, 21: 117-139.